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Aerosol organic carbon to black carbon ratios: Analysis of published data and implications for climate forcing

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Abstract

Measurements of organic carbon (OC) and black carbon (BC) concentrations over a variety of locations worldwide, have been analyzed to infer the spatial distributions of the ratios of OC to BC. Since these ratios determine the relative amounts of scattering and absorption, they are often used to estimate the radiative forcing due to aerosols. An artifact in the protocol for filter measurements of OC has led to widespread overestimates of the ratio of OC to BC in atmospheric aerosols. We developed a criterion to correct for this artifact and analyze corrected OC to BC ratios. The OC to BC ratios, ranging from 1.3 to 2.4, appear relatively constant and are generally unaffected by seasonality, sources or technology changes, at the locations considered here. The ratios compare well with emission inventories over Europe and China but are a factor of two lower in other regions. The reduced estimate for OC/BC in aerosols strengthens the argument that reduction of soot emissions maybe a useful approach to slow global warming.

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Aerosol organic carbon to black carbon ratios: Analysis of

published data and implications for climate forcing

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Abstract. Measurements of organic carbon (OC) and black carbon (BC) concentrations over a variety of locations worldwide, have been analyzed to infer the spatial distributions of the ratios of OC to BC. Since these ratios determine the relative amounts of scattering and absorption , they are often used to estimate the radiative forcing due to aerosols. An artifact in the protocol for filter measurements of OC has led to widespread overestimates of the ratio of OC to BC in atmospheric aerosols. We developed a criterion to correct for this artifact and analyze corrected OC to BC ratios. The OC to BC ratios, ranging from 1.3 to 2.4, appear relatively constant and are generally unaffected by seasonality, sources or technology changes, at the locations considered here. The ratios compare well with emission inventories over Europe and China but are a factor of two lower in other regions. The reduced estimate for OC/BC in aerosols strengthens the argument that reduction of soot emissions maybe a useful approach to slow global warming.

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1. Introduction

Carbonaceous aerosols produced by incomplete combustion of fossil and biomass fuels are ubiquitous in the atmosphere at concentrations comparable to sulfates, the major inorganic aerosol species. Sulfate aerosols influence climate through their light-scattering properties. Carbonaceous aerosols are composed of both light-absorbing black carbon (BC) and light-scattering organic carbon (OC). In this paper, we are concerned with the relative amounts of OC and BC in aerosol particles, particularly as it relates to the radiative forcing of climate.

Whether carbonaceous particles principally warm or cool the atmosphere depends on the aerosol single scattering albedo (SSA), which is the ratio of incident radiation that is scattered to the incident radiation that is absorbed. It follows that the aerosol SSA depends on the relative amounts of absorbing BC and scattering OC. An artifact in the filter measurements of OC has led to significant overestimates of the ratio of OC to BC in atmospheric aerosols. We developed a procedure to correct the published OC/BC ratios for this artifact. The corrected ratios are significantly lower than reported in the literature. This implies that carbonaceous particles may have lower OC/BC ratios. have lower SSA than previously assumed, and thus will more likely have a warming effect on climate.

The net effect of the radiative forcing of climate by carbonaceous aerosols is assessed by modeling studies. In climate models, predicted OC/BC ratios are based on OC and BC emissions estimated from sector segregated fuel consumption and relevant OC and BC emission factors [Cooke et al. 1999; Bond et al. 2004]. Estimated emission inventories of both OC and BC, and the derived OC/BC ratios, are subject to significant

uncertainties in both the fuel data and especially the adopted emission factors, signifying the need to compare the inventory-based and measured OC/BC ratios.

Measured OC and BC concentrations values, and therefore the OC/BC ratios, are also subject to uncertainties. These uncertainties are the result of sampling and analytical artifacts, which tend to either over- or under-estimate the concentrations of both OC and BC. OC concentrations are commonly determined by analysis of particles collected with quartz filters. During sample collection, gaseous organic compounds present in the sampled air adsorb onto the filter material. This process, known as the positive artifact, increases the OC of the filter deposit and thus the OC/BC ratio [Kirchstetter et al. 2001 and references therein]. Thus, measured OC emissions, as well as the OC/BC ratio, will be artificially high if the positive OC sampling artifact is overlooked. For example, uncorrected measurements during ACE-Asia project overestimated particle-phase OC by 20 to 100% [Mader et al., 2003]. Measured OC/BC ratios in biomass smoke plumes in Southern Africa were nearly twice as large if OC was not corrected for the positive artifact [Kirchstetter et al., 2003].

Additionally, some of the volatile components of the particle phase OC may be desorbed from the filter during sampling. This process is known as the negative OC sampling artifact and results in an underestimation of OC concentrations [*Eatough et al.* 1993].

The main uncertainty in introduced by the analytical methods pertains to their ability to accurately separate BC and OC [Schmid et al. 2001] and, therefore, bias the published OC and BC values. The repercussions of these analytical problems on our findings are discussed in the subsequent sections of this paper.

Another uncertainty relates to the fact that commonly used analytical methods quantify aerosol OC as the mass of carbon (generally expressed in units of µg C m⁻³) content of the organic aerosol material (OM). The mass of (OM) is larger than OC because OM contains other species besides carbon. As the OM is the relevant quantity for climate studies its concentration is often approximated by OC concentrations multiplied by a constant factor. Climate models, for example, commonly assume that OM is 1.3 or 1.4 times greater than the OC. The value of the factor relating OM and OC, however, is not accurately known [*Turpin and Lim* 2001]. In the following we consider OC, since it is the quantity that is actually measured.

Here we present an analysis of OC/BC ratios derived from published OC and BC concentrations measured in mostly urban and some non-urban locations in Asia, Europe and North America. Most of these data have not been corrected for the positive OC sampling artifact. Therefore, in the following discussion we refer to these as "apparent" concentrations and ratios. Accordingly, after presenting the published data, we describe a procedure to distinguish between data that is positive artifact-affected and positive artifact unaffected based on the nature of the OC sampling artifact. We then use unaffected data to infer regional and seasonal OC/BC ratios and make comparisons with values of OC/BC based on emission inventories.

2. Results

The OC and BC concentration data used in this study were obtained from filter samples analyzed by a number of investigators using different analytical methods. These data are taken from the literature and arranged roughly by year of publication (Table 1).

They include long term (i.e., annual) and short term (i.e., daily and weekly) averaged concentrations measured at sites affected by different fuel types, fuel consumption, and combustion technologies. Table 1 also indicates the analytical methods and aerosol size cuts employed in sample collection. Analytical methods denoted as "TO" combine thermal an optical measurement approach to improve OC – BC separation. "T" stands for solely thermal, usually temperature programmed analysis. Two-step methods "2ST" relay on exposing the sample to a fixed temperature, chosen to remove the OC from the sample. Carbon content of the exposed sample is operationally defined as BC, and the difference between the total carbon of unexposed and exposed sample gives the OC. "MNO" method uses MnO₂ as the oxygen donor for carbon oxidation at several pre-set temperatures. The details of the methods are in the references listed in Table 1.

To compare objectively the OC/BC ratios and BC concentrations, the relative standard deviations (rsd = 100% * standard deviation/average) of OC, BC and OC/BC are included in Table 1 for each region considered.

2.1. Spatial and seasonal OC/BC variations

OC/BC ratios for sites in China (Fig. 1a) are nearly constant (average 2.39 ± 0.47) irrespective of large variations in the range of BC concentration (4 µg BC m⁻³ to 15 µg BC m⁻³). OC/BC ratios appear to be the same in Beijing and Shanghai (2.39 ± 0.32), Pearl River Delta sites (2.44 ± 0.28) and Hong Kong (2.28 ± 0.88).

OC/BC ratios in Japan are similar in urban, coastal and island sites (Fig. 1b). This insensitivity to location and BC concentrations is qualitatively similar to that observed in China (Fig. 1a). However, the average OC/BC value for Japan (1.10 ± 0.31) is lower than measured in any other country we studied. The low value of this ratio in Japan suggests

that BC and OC emissions may be primarily derived from automotive sources, for which the OC/BC ratios are known to be low. For example, OC/BC ratios of automotive emissions in California from diesel and gasoline engines are 0.5 ± 0.4 and 0.9 ± 0.4 , respectively [*Kirchstetter et al.* 2004]. The OC/BC ratio measured in a tunnel in Austria is similar [*Leschober et al.* 2004].

OC/BC ratios for Korea and other Asian sites are distinctly different at continental and coastal sites (Fig. 1c). For Korean mainland sites the average value of OC/BC (1.51 \pm 0.46) falls between those for China and Japan. The OC/BC ratios for Korean island and coastal locations, however, are considerably higher. Average OC/BC ratio for Pakistan, and Mongolia (\approx 5) is substantially higher than the average for several cities in India and Bangladesh (\approx 1.9).

Plots of OC/BC distributions for European and North American locations are shown in Figure 2. European data (Fig. 2a) have an average OC/BC of 2.69 ± 0.91 for urban locations, increasing to values as high as 25 for regional background locations in Hungary, Italy and Sweden, where BC concentrations are low. The average OC/BC ratio (2.94 ± 1.61) for North American locations (Fig. 2b) is similar to the European ratios.

Finally, we note that BC concentrations at urban locations, such as Beijing, Shanghai, Chongju, Sapporo and Basel, show a significant increase in winter. However, no corresponding seasonal change in OC/BC ratios is observed at these locations (Table 2). At a few other locations such as Cheju Island and Gif sur Yvette (Table 1) there might be indications of seasonal differences in the OC/BC ratios. At these sites, however, seasonal BC variations are less pronounced and BC concentrations are much lower than in the examples shown in Table 2.

2.2 OC/BC ratios and BC concentrations

As indicated above, the apparent OC/BC ratios for China and Japan (Fig. 1a, b) do not depend appreciably on BC concentration. However, the OC/BC ratios for Korea (Fig. 1c), Europe, and North America, (Fig. 2a, b) are highly variable and appear to depend on BC concentration. High ratios are systematically associated with the lowest BC concentrations, and low ratios with the highest BC concentrations.

The OC/BC vs. BC plots for Asia, Europe, North America and other locations illustrate this trend. A plot of all Asian data (Fig 3a) shows that high ratios (mostly for coastal Korea) are clustered at BC concentrations < 1 μ g m⁻³. Low ratios for Japan (\approx 1.0), urban Korea (\approx 1.5), all of China (\approx 2) and other Asian locations correspond to high BC concentrations. European data (Fig. 3b) show similar features: extremely high ratios (μ to \approx 25) for BC < 1 μ g m⁻³ at "background" locations in Italy, Hungary and Sweden, and ratios of 2.41 \pm 0.86 for BC concentrations > 1.5 μ g m⁻³. In North America (Fig. 3c) half of all data points that show considerable scatter and relatively large ratios are for BC < 2 μ g m⁻³. The average OC/BC ratio for BC concentrations above 2 μ g m⁻³ is 1.88 \pm 0.64. Average OC/BC for all data considered here (Table 1) is 3.32 \pm 3.48, whereas this ratio drops to 2.20 \pm 1.51 for BC concentrations >1.5 μ g m⁻³.

3. Discussion

3.1. Correction criteria.

The results presented above demonstrate that high apparent OC/BC values systematically occur only with low BC concentrations. Next we show that neglect of the

positive sampling artifact, which causes an overestimation of OC, is responsible for the observed OC/BC vs. BC relationship.

The dependence of the apparent OC/BC ratios on BC concentrations is expected from the positive artifact mechanism. The magnitude of this artifact is related to the limited capacity of the filter for retaining the adsorbed (artifact) gas-phase OC. As a consequence the loading of adsorbed OC becomes a smaller fraction of total OC when the amount of gaseous (and particulate) species passing through the filter is high [Kirchstetter et al. 2001; Lunden et al., in preparation].

Positive artifact corrected OC concentration can be obtained by sampling with two quartz filters placed back-to-back as described by *Kirchstetter et al.* [2001] and *Mader et al.* [2003]. The top, or front, filter removes all particles, while gas-phase organics adsorb onto both the front and back filters. Thus, the measured carbon content of the backup filter can be subtracted from that of the front filter to give an estimate of the particle phase OC. (Uncorrected or apparent OC/BC is derived from the front filter only.)

That the observed high OC/BC ratios at low BC concentrations are greatly overestimated is demonstrated by comparing the uncorrected ("apparent") and positive artifact-corrected OC/BC ratios. During the SAFARI 2000 project [Swap et al. 2003], OC and BC concentrations were determined for a large number of aircraft collected samples taken in and out of biomass burning plumes [Kirchstetter et al. 2003]. The OC concentrations reported in that study were corrected for artifact OC by the method mentioned above. Fig. 4 shows the different dependence of artifact-corrected and uncorrected ratios on BC concentration. The uncorrected SAFARI ratios are as high as 27

at the lowest BC concentrations, and asymptotically approach the corrected value of about 6 at BC concentrations $> 1 \, \mu g \, m^{-3}$.

Artifacts inherent in some analytical methods may, in addition to positive and negative sampling artifacts, cause erroneous OC/BC ratios. Thermal analysis methods rely on heating a sample in an oxidizing or inert atmosphere and measuring the gaseous species evolved from the sample as a function of the sample temperature. This enables a separation of the carbonaceous material according to their volatilization, decomposition and combustion characteristics. The component that evolves at the highest temperature in then operationally defined as the black or elemental carbon. However, if a char-like material has been formed by OC pyrolysis then this fraction may be erroneously analyzed as the BC and thus cause a lower OC/BC value. One way to correct for the charring artifact is to perform optical transmission or reflectance measurement on the sample simultaneously with thermal analysis. Monitoring the optical signal allows for following the charring process and to more accurately quantify the actual BC and OC content and the OC/BC ratio.

We can evaluate the effect of the charring artifact on OC/BC ratios by considering a set of data (from Table 1) for Texas, Tennessee and Mexico City [*Fraser et al.* 2002; *Tanner et al.* 2004; *Chow et al.* 2002]. All of these data were obtained by the same analytical method Thermal Optical Reflection (TOR) [*Chow et al.* 1993], that corrects for the charring effect. An examination of this data set shows that the for low BC concentrations ($< 2 \mu g m^{-3}$) the average OC/BC is 4.22 ± 1.42 . At higher BC concentrations ($>2 \mu g m^{-3}$), however, the average OC/BC is 1.90 ± 0.64 . This trend is consistent with the trend of apparent OC/BC ratios, such as shown in Fig. 3, suggesting

that accounting for the charring artifact does not alter our positive sampling artifact based interpretation of the OC/BC vs. BC trend.

Lastly, we note that the data considered are insufficient to evaluate the effect of negative sampling artifact on OC/BC ratios.

3.2. Corrected OC/BC ratios. Based on the foregoing discussion we can identify apparent OC/BC ratios that can be used as "substitutes" for the artifact-corrected values. These unaffected ratios do not significantly depend on BC concentration variations. For example, all ratios for China and Japan (Fig. 1a, b) satisfy this criterion. For the European (Fig.2a), North American (Fig. 2b) and composite Asian data (Fig. 1c), we assume that the approximately constant OC/BC ratios, for BC concentrations > 1.5 μ g m⁻³ - 2.0 μ g m⁻³, represent the range of "correct" ratios.

The apparent and corrected OC/BC ratios together with corresponding average BC concentrations (in μg m⁻³) are summarized in Table 3. These data show that the unaffected (uncorrected) ratios can be lower than the apparent ratios by as much as a factor of two.

A more quantitative evaluation of the relative constancy of OC/BC compared to BC, is provided by considering the relative standard deviations (rsd) of OC, BC and OC/BC shown in Table 1. The rsd of the OC/BC ratio is about half of the rsd of BC (and OC) concentration for China, Japan, and Korea (and other Asian sites), supporting the conclusions given above. By contrast, the rsd of the OC/BC ratio is substantially larger than that of BC (and OC) for Europe, showing that OC/BC ratios are more variable than BC (and OC) in Europe. However, as asserted above, European OC/BC data are significantly overestimated when BC < $1.5 \mu g m^{-3}$. Limiting the analysis to data from

Europe where BC >1.5 ugm⁻³ reveals that the RSD of the OC/BC ratio is about half that of BC, as observed for China, Japan and Korea.

3.3. Comparison with results from emission factors

It is of interest to compare our derived ratios with OC and BC emission ratios calculated from emission factors and fuel consumption. The fuel and technology data used in such calculations are for entire countries, not for particular urban and non-urban locations where the measurements listed in this paper were made. We note, however, that the derived OC/BC ratios examined in this study do not show significant site-to-site and seasonal variations within a country, suggesting that average OC/BC approximately reflect the countrywide emissions. Furthermore, as the majority of these data (Table 1) were obtained in source dominated (urban) sites we may assume that the OC and BC at these sites are primary emission species.

Table 4 compares calculated and measured (corrected) regional OC/BC ratios. Calculated ratios are based on OC and BC emissions (for 1996) separated into contained, fossil and biofuel, and open biomass burning [Bond et al. 2004]. As the data in Table 4 show, the measured and emission inventory based total OC/BC ratios agree for Europe and China. Measured values, however, are significantly lower for North America and India. Total emission inventory based ratio for "Other Asian Countries" is more than twice the measured values for the three Asian countries and for INDOEX. These ratios are similar to the fossil fuel ratios, consistent with previous estimates of relative contributions of fossil and biomass sources for the Indian Ocean region [Mayol-Bracero et al. 2002; Novakov et al. 2000].

3.4. OC/BC ratios and radiative forcing. The relevance of artifact-corrected OC/BC ratios to the radiative forcing of climate by aerosols can be obtained by comparing them to those used in climate models. The carbonaceous aerosols in the models are commonly obtained either via a specification of the OC/BC ratio (e.g., Koch [2001] assumed OC/BC was 4 for fossil fuels and 8 for biomass sources) or via separate specification of OC and BC emission distributions. In either case, the resulting global distributions of carbonaceous aerosols will tend to overestimate the proportion of OC if the input data is not corrected for the positive sampling artifact. We compare in Fig. 5, uncorrected values of observed OC/BC with the OC/BC ratios obtained in a current model, the Goddard Institute for Space Studies (GISS) ModelE [Schmid et al. 2005]. The spatial distributions of the OC and BC emissions are based on Bond et al. [2004] for fossil fuel and bio fuel sources and on Van der Werf et al. [2003] for biomass sources. The OC/BC ratios derived from this model (in a simulation that includes both the direct and indirect aerosol effects from *Menon and Del Genio* [2005]) are compared to observed values in Fig. 5a. In general, the model over estimates the ratios, with higher ratios predicted over most parts of the U.S., India, parts of Asia and over the Mediterranean with closer agreement (within 20%) mainly over China (Fig. 5b).

4. Conclusions

Our analysis of measured ambient OC/BC mass concentration ratios demonstrates that the neglect of a positive sampling artifact for OC results in a large overestimation of the OC/BC ratios, especially at low BC concentration. A survey of relevant publications (Table 1) shows that most data need to be corrected for this artifact. We have developed

criteria to select those uncorrected OC/BC ratios that can be used as reasonable approximations for correct data. These criteria are defined using the results of studies in which both corrected and uncorrected ratios were determined.

After artifact correction, the OC/BC ratios exhibit remarkable constancy for specific regions. Overall the ratios range from about 1.3 in Japan to about 2.4 at other locations. The OC/BC ratio shows no significant seasonal variability and appears to be insensitive to regional fuel and technology mix.

Comparison of our empirical regional ratios with those calculated from emission inventories for both fossil and biomass sources [*Bond et al.* 2004] show good agreement for China and Europe. Our ratios, however, are about a factor of 2 lower than the values calculated from OC and BC emission inventories for all other regions.

The generally reduced values of OC/BC that we find after correction for the positive OC artifact has relevance to the climate effect of anthropogenic soot aerosols. It has been suggested [Hansen et al. 2000; Jacobson, 2001] that reduction of anthropogenic BC emissions would help slow global warming. Penner et al. [2003], on the other hand, have pointed out that the cooling effect of OC that inevitably accompanies BC, together with the indirect effects of both BC and OC on cloud properties, make it uncertain whether the net effect of soot (BC + OC) emissions is warming or cooling. Hansen et al. [2005] calculate that the net effect of fossil fuel soot is warming, while biomass burning produces global cooling. The effect of reduced estimates for the OC/BC ratio, given in this paper, is to tilt the calculations for the effect of soot on climate more toward warming. The most effective targets for soot emission reductions would be those with especially low values of OC/BC, such as vehicles using diesel fuel. Reduction of aerosol

emissions from such sources are most likely to reduce global warming as well improve public health [Schneider and Hill, 2005].

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References

- Bond, T.C., D.G. Street, K.F. Yarber, S.M. Nelson, J._H. Woo, and Z. Klimont (2004), A technology-based global inventory of black and organic carbon emissions from combustion, J. *Geophys. Res.*, *109*, D14203, doi: 10/1029/2003JD003697.
- Cachier, H., M-P. Brémond, and P. Buat-Ménard (1989), Thermal separation of soot carbon, *Aerosol Sci. Technol.*, *10*, 358-364.
- Cadle, S.H., and J.M. Dasch (1988), Wintertime concentrations and sinks of atmospheric particulate carbon at a rural location in northern Michigan, *Atmos. Environ.*, 22, 1373-1381.
- Cao, J.J., S.C. Lee, K.F. Ho, X.Y. Zhang, S.C. Zou, K. Fung, J. C. Chow, and J. G.Watson (2003), Characteristics of carbonaceous aerosol in Pearl River DeltaRegion, China during 2001 winter period, *Atmos. Environ.*, 37, 1451–1460.
- Carvalho, A., C. Pio, and C. Santos (2003), Water-soluble hydroxylated organic compounds in German and Finnish aerosols, *Atmos. Environ.*, *37*, 1775–1783.
- Castro, L.M., C.A. Pio, R.M. Harrison, and D.J.T. Smith (1999), Carbonaceous aerosols in urban and rural European atmospheres: estimation of secondary organic carbon concentrations, *Atmos. Environ.*, *31*, 2771-2781.
- Chen, L.-W. A., B. G. Doddridge, R. R. Dickerson, J. C. Chow, and R. C. Henry (2002), Origins of fine aerosol mass in the Baltimore–Washington corridor: implications from observation, factor analysis, and ensemble air parcel back trajectories, *Atmos. Environ.*, 36, 4541–4554.

- Chow, J. C., J. G. Watson, S. A. Edgerton, and E. Vega (2002), Chemical composition of PM 2.5 and PM 10 in Mexico City during winter 1997, *Sci. Total Environ.*, 287, 177-201.
- Chow, J.C., J. G. Watson, L.C. Pritchett, W.R. Pierson, C.A. Frazier, and R. G. Purcell (1993), The DRI thermal/optical reflectance carbon analysis system: description, evaluation and application in U.S. air quality studies, *Atmos. Environ.*, *27A*, 1185 1201.
- Chow, J.C., J. G. Watson, D.J. Lowenthal, P.A. Solomon, K.L. Magliano, S.S. Ziman, and L.W. Richards (1993), PM₁₀ and PM_{2.5} compositions in California's San Joaquin Valley, *Aerosol Sci. Technol.*, *18*, 105-128.
- Cooke, W. F., C. Liousse, H. Cachier, and J. Feichter (1999), Construction of a 1° × 1° fossil fuel emission data set for carbonaceous aerosol and implementation in the ECHAM4 model (1999), *J. Geophys. Res.*, 104 (D18), 22,137–22,162.
- Countess, R.J., S.H. Cadle, P.J. Groblicki, G.T. Wolff (1981), J. Air Poll. Contr. Assoc., 31, 247-252.
- Dan, M., G. Zhuang, X. Li, H. Tao, Y. Zhuang (2004), The characteristics of carbonaceous species and their sources in PM2.5 in Beijing, *Atmos. Environ.*, 38, 3443–3452.
- Decesari, S., M. C. Facchini, E. Matta, F. Lettini, M. Mircea, S. Fuzzi, E. Tagliavini and J. -P. Putaud (2001), Chemical features and seasonal variation of fine aerosol water-soluble organic compounds in the Po Valley, Italy, *Atmos. Environ.*, *35*, 3691-3699.
- Eatough, D. J., F. Obeidi, Y. Pang, Y. Ding, N.L. Eatough, and W.E. Wilson (1999),

- Integrated and real-time diffusion denuder sampler for $PM_{2.5}$, *Atmos. Environ.*, 33, 2835 2844.
- Fraser, M.P., Z.W. Yue, R.J. Tropp, S.D. Kohl, and J.C. Chow (2002), Molecular composition of organic fine particulate matter in Houston, TX, *Atmos. Environ.*, *36*, 5751–5758.
- Hansen, J.E., M. Sato, R. Ruedy, A. Lacis, and V. Oinas (2000), Global warming in the twenty-first century: an alternative scenario, *Proc. Natl. Acad. Sci.*, **97**, 9875-9880.
- Hansen, J.E., M. Sato, R. Ruedy, L. Nazarenko, A. Lacis, G.A. Schmidt, G. Russell, I.
 Aleinov, M. Bauer, S. Bauer, N. Bell, B. Cairns, V. Canuto, M. Chandler, Y.
 Cheng, A. Del Genio, G. Faluvegi, E. Fleming A. Friend, T. Hall, C. Jackman, M.
 Kelley, N. Kiang, D. Koch, J. Lean, J. Lerner, K. Lo, S. Menon, R. Miller, P.
 Minnis, T. Novakov, V. Oinas, Ja. Perlwitz, Ju. Perlwitz, D. Rind, D. Romanou,
 D. Shindell, P. Stone, S. Sun, N. Tausnev, D. Thresher, B. Wielicki, T. Wong, M.
 Yao, and S. Zhang (2005) Efficacy of climate forcings, *J. Geophys. Res.*, In press.
- Hatakeyama, S. (1993), Data of '91 IGAC/APARE/PEACAMPOT Survey. National Institute for Environmental Studies, Japan.
- He, K., F. Yang, Y. Ma, Q. Zhang, X. Yao, C. K. Chan, S. Cadle, T. Chan, and P. Mulawa (2001), The characteristics of PM2.5 in Beijing, China, *Atmos. Environ.*, *35*, 4959–4970.
- He, Z., Y.J. Kim, K.O. Ogunjobi, J.E. Kim, and S.Y. Ryu (2004), Carbonaceous aerosol characteristics of PM 2.5 particles in Northeastern Asia in summer 2002, *Atmos. Environ.*, 38, 1795–1800.

- Ho, K.F., S.C. Lee, C. K. Chan, J. C. Yu, J. C. Chow, and X.H. Yao (2003),

 Characterization of chemical species in PM2:5 and PM 10 aerosols in Hong Kong, *Atmos. Environ.*, *37*, 31–39.
- Holler, R., S. Tohno, M. Kasahara, and R. Hitzenberger (2002), Long-term characterization of carbonaceous aerosol in Uji, Japan, *Atmos. Environ.*, *36*, 1267–1275.
- Jacobson, M.Z., 2001: Strong radiative heating due to the mixing state of black carbon in atmospheric aerosols. *Nature*, 409, 695-697.
- Kadowaki, S. (1990), Characterization of carbonaceous aerosols in Nagoya urban area 1.
 Elemental and organic carbon concentrations and the origin of organic aerosols,
 Envir. Sci. Technol. 24, 741-744.
- Kaneyasu, N., S. Ohta, and N. Murao (1995), Seasonal variation in the chemical composition of atmospheric aerosols and gaseous species in Sapporo, Japan, *Atmos. Environ.*, 29, 1559-1568.
- Keeler, G.J., J.D. Spengler, P. Koutrakis, and G. A. Allen (1990), Transported acid aerosols measured in southern Ontario, *Atmos. Environ.*, *24A*, 2935-2950.
- Keeler, G. J., S.M. Japar, W.W. Brachaczek, R.A. Gorse, Jr., J.M. Norbeck, and W.R.
 Pierson (1986), *Atmos. Environ.*, 20, 1281-1289. Kendall, M., R.S. Hamilton, J.
 Watt, and I.D. Williams (2001), Characterisation of selected speciated organic compounds associated with particulate matter in London, *Atmos. Environ.*, 35, 2483-2495.

- Kim, Y. P., K-C. Moon, J. H. Lee, N. J. Baik (1999), Concentrations of carbonaceous species in particles at Seoul and Cheju, Korea during 1994, *Atmos. Environ.*, *33*, 2751-2758.
- Kim, Y.P., J.H. Lee, H.J. Baik, J. Y. Kim, S-G. Shim, and C-H. Kang (1998),
 Summertime characteristics of aerosol composition at Cheju Island, Korea,
 Atmos. Environ., 32, 3905-3915.
- Kirchstetter, T.W., C.E. Corrigan, and T. Novakov (2001), Laboratory and field investigation of the adsorption of gaseous organic compounds onto quartz filters, *Atmos. Environ.*, *35*, 1663-1671.
- Kirchstetter, T.W., T. Novakov, P. V. Hobbs, and B. Magi (2003), Airborne measurements of carbonaceous aerosol in southern Africa during the dry biomass burning season, *J. Geophys. Res.*, 108 (D13), 8476, doi:10.1029/2002JD002171.
- Kirchstetter, T., D. Hooper, Z. Apte, A. Strawa, G. Hallar, R. Harley, G. Kendall, J. Hesson, E. Stevenson, A. Miguel, and A. Eiguren-Fernandez (2004),
 Characterization of particle and gas phase pollutant emissions from heavy- and light-duty vehicles in a California roadway tunnel, Paper presented at AGU Fall Meeting, San Francisco, December 2004.
- Koch, D, 2001: Transport and direct radiative forcing of carbonaceous and sulfate aerosols in the GISS GCM. J. Geophys. Res., 106, 20311-20332.
- Laschober, C., A. Limbeck, J. Rendl, and H. Puxbaum (2004), Particulate emissions from on-road vehicles in the Kaisermühlen-tunnel (Vienna, Austria), *Atmos. Environ.*, 38, 2187–2195.

- Lee, H.S., and B-W. Kang (2001), Chemical characteristics of principal PM2.5 species in Chongju, South Korea, *Atmos. Environ.*, *35*, 739-746.
- Lee, J.H., Y.P. Kim, K.-C. Moon, H.-K. Kim, and C.B. Lee (2001), Fine particle measurements at two background sites in Korea between 1996 an 1997, *Atmos. Environ.*, *35*, 635-643, 2001.
- Mader, B. T., J. J. Schauer, J. H. Seinfeld, R. C. Flagan, J. Z. Yu, H. Yang, Ho-Jin Lim,
 B. J. Turpin, J. T. Deminter, and G. Heidemann (2003), Sampling methods used
 for the collection of particle-phase organic and elemental carbon during ACEAsia, Atmospheric Environ., 37, 1435-1449.
- Mayol-Bracero, O.L., R. Gabriel, M. O. Andreae, T. W. Kirchstetter, T. Novakov, J.
 Ogren, P. Sheridan, and D. G. Streets (2002), Carbonaceous aerosols over the
 Indian Ocean during the Indian Ocean Experiment (INDOEX): Chemical
 characterization, optical properties, and probable sources, *J. Geophys. Res.*, 107
 (D19), 8030, doi:10.1029/2000JD000039.
- Menon, S., and A.D. Del Genio (2005), Evaluating the Impacts of Carbonaceous

 Aerosols on Clouds and Climate, In "Human-Induced Climate Change: An

 Interdisciplinary Assessment", M. Schlesinger et al., Eds (Cambridge Univ. Press,

 New York).
- Molnár, A., E. Mészáros, H.C. Hansson, H. Karlsson, A. Gelencsér, Gy. Kiss, and Z. Krivácsy (1999), The importance of organic and elemental carbon in the fine atmospheric aerosol particles, *Atmos. Environ.*, *33*, 2745-2750.
- Muhlbaier-Dasch, J., and S.H. Cadle (1989), Atmospheric carbon particles in the Detroit urban area: Wintertime sources and sinks, *Aerosol Sci. Technol.*, 10, 236-248.

- Novakov, T., M.O. Andreae, R. Gabriel, T.W. Kirchstetter, O.L. Mayol-Bracero, and V. Ramanathen (2000), Origin of carbonaceous aerosols over the tropical Indian Ocean: Biomass burning or fossil fuels?, *Geophys. Res. Letters*, *27*, 4061-4064.
- Ohta, S., and T. Okita (1984), Measurements of particulate carbon in urban and marine air in Japanese areas, *Atmos. Environ.*, 18, 2439-2445.
- Ohta, S., M. Hori, S. Yamagata, and N. Murao (1998), Chemical characterization of atmospheric fine particles in Sapporo with determination of water content, *Atmos. Environ.*, 32, 1021-1025.
- Park, S.S., Y.J. Kim, and K. Fung (2001), Characteristics of PM 2.5 carbonaceous aerosols in the Sihwa industrial area, South Korea, *Atmos. Environ.*, *35*, 657-665.
- Park, S. S., Y. J. Kim, and K. Fung (2002), PM2.5 carbon measurements in two urban areas: Seoul and Kwangju, Korea, *Atmos. Environ.*, *36*, 1287–1297.
- Penner, J.E., S.Y. Zhang, and C.C. Chuang (2003), Soot and smoke may not warm climate, *J. Geophys. Res.*, 108 (D21), 4657, doi:10.1029/2003JD003409.
- Pio, C.A., L.M. Castro, M.A. Cerqueira, I.M. Santos, F. Belchior, and M. L. Salgueiro (1996), Source assessment of particulate air pollutants measured at the southwest European coast, *Atmos. Environ.*, *30*, 3309-3320.
- Röösli, M., G. Theis, N. Künzli, J. Staehelin, P. Mathys, L. Oglesby, M. Camenzind, and Ch. Braun-Fahrländer (2001), Temporal and spatial variation of the chemical composition of PM10 at urban and rural sites in the Basel area, Switzerland, *Atmos. Environ.*, *35*, 3701-3713.
- Saitoh, K., K. Sera, K. Hirano, and T. Shirai (2002), Chemical characterization of particles in winter-night smog in Tokyo, *Atmos. Environ.*, *36*, 435–440.

- Salam, A., H. Bauer, K. Kassin, S. M. Ullah, and H. Puxbaum (2003a), Aerosol chemical characteristics of an island site in the Bay of Bengal (Bhola Bangladesh), *J. Environ. Monit.*, *5*, 1-9.
- Salam, A., H. Bauer, K. Kassin, S. M. Ullah, and H. Puxbaum (2003b), Aerosol chemical characteristics of a mega-city in Southeast Asia (Dhaka–Bangladesh), *Atmos. Environ.*, *37*, 2517–2528.
- Salma, I., X. Chib, and W. Maenhaut (2004), Elemental and organic carbon in urban canyon and background environments in Budapest, Hungary, *Atmos. Environ.*, 38, 27–36.
- Shah, J.J., T.J. Kneip, and J.M. Daisey (1985), Source apportionment of carbonaceous aerosol in New York City by multiple linear regression, J. Air Poll. Contr. Assoc., *35*, 541-544.
- Schneider, C.G. and L.B. Hill (2005), Diesel and health in America: The lingering threat.

 Clean Air Task Force Report, Edited by M. Padian, Spectrum Printing and

 Graphics, Inc., Boston, MA.
- Schmid, H., L. Laskus, H. J. Abraham, U. Baltensperger, V. Lavanchy, M. Bizjak, P.
 Burba, H. Cachier, D. Crow, J. Chow, T. Gnauk, A. Even, H.M. ten Brink, K.P. Giesen, R. Hitzenberger, C. Hueglin, W. Maenhaut, C. Pio, A. Carvalho, J.- P.
 Putaud, D. Toom-Sauntry, and H. Puxbaum (2001), Results of the carbon
 conference international aerosol carbon round robin test stage I, *Atmos. Environ.*,
 35, 2111 2121.

- Schmidt, G. A., R. Ruedy, J.E. Hansen, I. Aleinov, N. Bell, M. Bauer, et al. (2005),

 Present day atmospheric simulations using GISS ModelE: Comparison to in-situ, satellite and reanalysis data, *J. Clim.*, In Review.
- Smith, D.J.T., R.M. Harrison, L. Luhana, C.A. Pio, L.M. Castro, M.N. Tario, S. Hayat, and T. Quraishi (1996), Concentrations of particulate airborne polycyclic aromatic hydrocarbons and metals collected in Lahore, Pakistan, *Atmos. Environ.*, 30, 4031-4040.
- Solomon, P.A., and J. L. Moyers (1986), A chemical characterization of wintertime haze in Phoenix, Arizona, *Atmos. Environ.*, 20, 207-213.
- Solomon, P.A., T. Fall, L. Salmon, G.R. Cass, H.A. Gray, and A. Davidson (1989),

 Chemical characteristics of PM₁₀ aerosol in the Los Angeles area, *J. Air. Poll.*Contr. Assoc., 39, 254-163.
- Swap, R. J., H.J. Annegarn, J. T. Suttles, M.D.King. S. Platnick, J. L. Privette, and R. J. Scholes (2003), Africa burning: A thematic analysis of the Southern African Regional Science Initiative (SAFARI 2000), *J. Geophys. Res.*, 108(D13), 8465, doi:10.1029/2003JD003747.
- Tanner, R. L., W. J.Parkhurst, M. L.Valente, and W.D. Phillips (2004), Regional composition of PM2.5 aerosols measured at urban, rural and "background" sites in the Tennessee valley, *Atmos. Environ.* 38, 3143–3153.
- Turpin, B. J. and H. J. Lim, (2001), Species contributions to PM2.5 mass concentrations: revisiting common assumptions for estimating organic mass, *Aerosol. Sci. Technol.*, 35, 602-610.

- Valaoras, G., J.J. Huntzicker, and W.H. White (1988), On the contribution of motor vehicles to the Athenian "nephos": An application of factor signatures, *Atmos. Environ.*, 22, 965-971.
- Van der Werf, G.R., J.T. Randerson, G.J. Collatz, L. Giglio, P.S. Kasibhatla, A.F.
 Arellano, S.C. Olsen, and E.S. Kasischke (2004), Continental-scale partitioning of fire emissions during the 1997 to 2001 El Nino/La Nina period, *Science*, 303, 73-76.
- Venkataraman, C., C. K. Reddy, S. Josson, and M.S. Reddy (2002), Aerosol size and chemical characteristics at Mumbai, India, during the INDOEX-IFP (1999), *Atmos. Environ.*, *36*, 1979–1991.
- Viidanojaa, J., M. Sillanpää, J. Laakia, V.-M. Kerminen, R. Hillamo, P. Aarnio, and T. Koskentalo (2002), Organic and black carbon in PM2.5 and PM10: 1 year of data from an urban site in Helsinki, Finland, *Atmos. Environ.*, *36*, 3183-3193.
- Wolff, G.T., and P.A. Korsog (1985), Estimates of the contributions of sources to inhalable particulate concentrations in Detroit, *Atmos. Environ.*, *19*, 1399-1409.
- Wolff, G.T., N.A. Kelly, M.A. Ferman, M.S. Ruthkosky, D.P. Stroup, and P.E. Korsog (1986), Measurements of sulfur oxides, nitrogen oxides, haze and fine particles at a rural site on the Atlantic coast, *J. Air. Poll. Contr. Asoc.*, *36*, 585-591.
- Wolff, G.T., P.E. Korsog, N.A. Kelly, and M.A. Ferman (1985), Relationship between fine particulate species, gaseous pollutants and meteorological parameters in Detroit, *Atmos. Environ.*, 19, 1341-1349.

- Wu, Y., J. Hao, L. Fu, J. Hu, Z. Wang, and U. Tang (2003), Chemical characteristics of airborne particulate matter near major roads and at background locations in Macao, China, Sci. Total Environ., 317, 159-172.
- Ye, B., X. Ji, H. Yang, X. Yao, C. K. Chan, S. H. Cadle, T. Chan, and P. A. Mulawa (2003), Concentration and chemical composition of PM2.5 in Shanghai for a 1-year period, *Atmos. Environ.*, *37*, 499–510.
- Yu, J. Z., J. W. T. Tung, A. W. M. Wu, A. K. H. Lau, P. K.-K. Louie, and J. C. H. Fung (2004), Abundance and seasonal characteristics of elemental and organic carbon in Hong Kong PM10, *Atmos. Environ.*, 38, 1511–1521.
- Zappoli, S., A. Andracchio, S. Fuzzi, M.C. Facchini, A. Gelencsér, G. Kiss, Z. Krivácsy,
 Á. Molnár, E. Mészáros, H.-C. Hansson, K. Rosman, and Y. Zebühr (1999),
 Inorganic, organic and macromolecular components of fine aerosol in different
 areas of Europe in relation to their water solubility, *Atmos. Environ.*, 33, 2722-2743.

Table captions

- Table 1. Data sources used in analysis. Designations and description of analytical methods given in text.
- Table 2. Seasonal changes in OC/BC ratios and BC concentrations (μg m⁻³)
- Table 3. Comparison of average BC concentrations and apparent and corrected OC/BC ratios
- Table 4. Calculated and corrected measured regional OC/BC ratios

Figure captions

- Fig. 1. Variations of BC concentrations and apparent OC/BC ratios for a) China, b) Japan, c) Korea and other Asia.
- Fig. 2. Same as in Fig. 1 but for a) Europe and b) North America.
- Fig. 3. Plots of apparent OC/BC ratios vs. BC concentrations for a) Asia, b) Europe and c) North America.
- Fig. 4. Comparison of uncorrected and positive artifact corrected OC/BC ratios measured during SAFARI 2000.
- Fig. 5. a) Observed OC/BC ratios and model (GISS GCM) predicted OC/BC ratios (based on annual averages) at the select locations where measurements were available; b) Differences (in percent) between measurements and model predicted OC/BC ratios.

Table 1. Data sources used in analysis. Designations and description of analytical methods given in text.

Location	Sampling dates	Particle size	OC	ВС	OC/BC	References
		cut, µm/	$\mu g m^{-3}$	$\mu g m^{-3}$		
		Analytical				
		method				
China						
Beijing 1	Summer 1999	2.5 / TO	13.42	6.27	2.14	He et al. [2001]
Beijing 2	Fall 1999	2.5 / TO	28.79	10.23	2.81	He et al. [2001]
Beijing 3	Winter 1999-2000	2.5 / TO	31.49	11.08	2.84	He et al. [2001]
Beijing 4	Spring 2000	2.5 / TO	18.21	6.67	2.73	He et al. [2001]
Beijing 5	Annual 1999-2000	2.5 / TO	25.30	9.40	2.69	He et al. [2001]
Shanghai 1	Hainan, summer 1999	2.5 / TO	13.07	5.71	2.29	Ye et al. [2003]
Shanghai 2	Hainan, fall 1999	2.5 / TO	17.28	6.97	2.48	Ye et al. [2003]
Shanghai 3	Hainan, winter 1999 - 20000	2.5 / TO	17.59	8.07	2.18	Ye et al. [2003]
Shanghai 4	Hainan, annual 1999-2000	2.5 / TO	15.98	6.92	2.31	Ye et al. [2003]
Shanghai 5	Tongji, spring 1999	2.5 / TO	16.10	5.27	3.06	Ye et al. [2003]
Shanghai 6	Tongji, summer 1999	2.5 / TO	9.62	4.61	2.09	Ye et al. [2003]
Shanghai 7	Tongji, fall 1999	2.5 / TO	15.22	6.81	2.23	Ye et al. [2003]
Shanghai 8	Tongji, winter 1999-2000	2.5 / TO	16.40	8.16	2.01	Ye et al. [2003]
Shanghai 9	Tongji, annual 1999-2000	2.5 / TO	14.34	6.21	2.31	Ye et al. [2003]
Hong Kong 1	PU site, Nov Feb. 2000 - 01	2.5 /MNO	9.45	5.80	1.63	<i>Ho et al.</i> [2003]

Hong Kong 2	KT site, Nov Feb. 2000 -01	2.5 / MNO	10.16	5.05	2.01	Ho et al. [2003]
Hong Kong 3	HT site Nov Feb. 2000 - 01	2.5 / MNO	5.52	1.36	4.06	Ho et al. [2003]
Hong Kong 4	Average 3 sites	2.5 / MNO	8.38	4.07	2.06	Ho et al. [2003]
Hong Kong 5	Jan Feb. 2002	2.5 / TO	9.60	4.70	2.04	Cao et al. [2003]
Shenzen	Jan Feb. 2002	2.5 / TO	13.20	6.10	2.16	Cao et al. [2003]
Guangzhou,	Jan- Feb, 2002	2.5 / TO	22.60	8.30	2.72	Cao et al. [2003]
Zhuhai	Jan Feb. 2002	2.5 / TO	12.20	5.00	2.44	Cao et al. [2003]
Pearl river delta	All sites Jan Feb. 2002	2.5 / TO	14.70	6.10	2.41	Cao et al. [2003]
Macao	Dec. July 2001, Dec. 2002	2.5 / TO	12.20	4.40	2.77	Wu et al. [2003]
Hong Kong 6	9 site average 1998-2001	10 / TO	8.89	4.66	1.91	Yu et al. [2004]
Beijing 6	Jun July 2002	2.5 / 2ST	10.70	5.70	1.88	Dan et al. [2004]
Beijing 7	Dec. 2002	2.5 / 2ST	36.70	15.20	2.41	Dan et al. [2004]
D.::: 0	S.,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,,	2.5./2CT	12.40	5.40	2.29	He at al [2004]
Beijing 8	Summer 2002	2.5 / 2ST	12.40	3.40	2.29	<i>He et al.</i> [2004]
Relative standard		2.5 / 281	47	40	20	не ei ai. [2004]
		2.3 / 281				не et at. [2004]
Relative standard		8.0 / 2ST				Ohta and Okita
Relative standard	deviation (%)		47	40	20	
Relative standard	deviation (%)		47	40	20	Ohta and Okita
Relative standard Japan Sapporo 1	Annual 1982	8.0 / 2ST	4.15	5.10	0.81	Ohta and Okita [1984]
Relative standard Japan Sapporo 1	Annual 1982	8.0 / 2ST	4.15	5.10	0.81	Ohta and Okita [1984] Ohta and Okita
Relative standard Japan Sapporo 1 Chichi-Jima	Annual 1982 Dec. 1981	8.0 / 2ST TSP / 2ST	4.15	5.10	0.81	Ohta and Okita [1984] Ohta and Okita [1984]
Relative standard Japan Sapporo 1 Chichi-Jima	Annual 1982 Dec. 1981	8.0 / 2ST TSP / 2ST	4.15	5.10	0.81	Ohta and Okita [1984] Ohta and Okita [1984] Ohta and Okita
Relative standard Japan Sapporo 1 Chichi-Jima Hachijo-Jima 1	Annual 1982 Dec. 1981 Jan. 1981	8.0 / 2ST TSP / 2ST	4.15 0.60 1.40	5.10 0.70 1.00	0.81 0.86	Ohta and Okita [1984] Ohta and Okita [1984] Ohta and Okita [1984]
Relative standard Japan Sapporo 1 Chichi-Jima Hachijo-Jima 1	Annual 1982 Dec. 1981 Jan. 1981	8.0 / 2ST TSP / 2ST	4.15 0.60 1.40	5.10 0.70 1.00	0.81 0.86	Ohta and Okita [1984] Ohta and Okita [1984] Ohta and Okita [1984] Ohta and Okita

						[1993]
Oki Island	1991	na	1.60	1.25	1.28	Hatekeyama et al.
						[1993]
Okinawa	1991	na	0.78	1.17	0.67	Hatekeyama et al.
						[1993]
Nagano	1991	na	1.20	1.50	0.80	Hatekeyama et al.
						[1993]
Sapporo 2	Site 1, Jun. 1987 - Dec 1988	10 /2ST	3.40	3.30	1.03	Kaneyasu et al.
						[1995]
Sapporo 3	Site 2, Jun. 1987 - Dec 1988	10 /2ST	3.60	3.60	1.00	Kaneyasu et al.
						[1995]
Sapporo 4	Annual 1991	10 / 2ST	3.74	4.26	0.88	Ohta et al. [1998]
Uji	Jan. 1989 - Nov. 1999	10 /RP	2.03	2.64	0.77	Höller et al.
						[2002]
Tokyo	Dec. 1998 - Jan. 1999	2.5 / RP	7.80	5.40	1.44	Saitoh et al.
						[2002]
Kyoto	Summer 2002	2.5 / 2ST	5.00	1.30	3.85	He et al. [2004]
Relative standard	deviation (%)		109	102	61	
Korea and other						
Asian countries						
Kosan 1	July 1994	2.5 / MNO	4.58	0.38	12.05	Kim et al. [1998]
Kosan 2	August 1994	2.5 / MNO	2.36	0.08	29.50	Kim et al. [1998]
Seoul 1	Jun. 1994	2.5 / MNO	9.97	7.57	1.32	Kim et al. [1999]
Chongju	Annual 1995-1996	2.5 / TO	4.99	4.44	1.12	Lee and Kang
						[2001]

Sihwa	1998- 1999	2.5 / MNO	9.1	1.8	5.06	Park et al. [2001]	
Kosan 3 (Cheju isl)	March 96	2.5 / MNO	2.97	0.32	9.28	Lee et al. [2001]	
Kosan 4 (Cheju isl)	Dec. 1996	2.5 / MNO	4.41	0.43	10.26	Lee et al. [2001]	
Kosan 5 (Cheju isl)	Jan. 1997	2.5 / MNO	3.31	0.23	14.39	Lee et al. [2001]	
Kosan 6 (Cheju isl)	Sep. 1997	2.5 / MNO	3.56	0.42	8.48	Lee et al. [2001]	
Kosan 7(Cheju isl)	Dec. 1997	2.5 / MNO	2.6	0.34	7.65	Lee et al. [2001]	
Kosan 8 Average	Annual average 1996 – 1997	2.5 / MNO	3.26	0.34	9.59	Lee et al. [2001]	
Kanghwa 1	March 1996	2.5 / MNO	5.16	0.56	9.21	Lee et al. [2001]	
Kanghwa 2	December 1996	2.5 / MNO	11	2.45	4.49	Lee et al. [2001]	
Kanghwa 3	Jan. 1997	2.5 / MNO	7.58	0.95	7.98	Lee et al. [2001]	
Kanghwa 4	Sep. 1997	2.5 / MNO	4.28	0.79	5.42	Lee et al. [2001]	
Kanghwa 5	Dec. 1997	2.5 / MNO	6.23	0.9	6.92	Lee et al. [2001]	
Kanghwa 6	Annual average 1996 – 1997	2.5 / MNO	6.45	0.98	6.58	Lee et al. [2001]	
Seoul 2	Nov - Dec. 1999	2.5 / MNO	15.2	7	2.17	Park et al. [2002]	
Kwangju	Jun. 2000	2.5 / MNO	7.6	5.3	1.43	Park et al. [2002]	
Gwangju	Summer 2002	2.5 / 2ST	1.4	0.3	4.67	He et al. [2004]	
Lahore	1992 - 1993	TSP/TO	74.7	17.53	4.26	Smith et al.,	
						[1996]	
Mumbai	Mar. 1999	10 / TO	25.3	12.6	2.01	Venkataraman et	
						al. [2002]	
Dhaka	Apr May 2001	TSP / 2ST	45.7	22	2.08	Salam et al.	
						[2003a]	
Bhola 1	May 14-17, 2001	TSP / 2ST	4.22	3.18	1.33	Salam et al.	
						[2003b]	
Bhola 2	May 1-21, 2001	TSP / 2ST	5.04	2.37	2.13	Salam et al.	

						[2003b]
Ulan Bator	July. 2002	2.5 / 2ST	2.3	0.4	5.75	He et al. [2004]
Relative standard deviation (%)			124	156	89	
Europe						
Athens 1	June – Aug. 1982	TSP / TO	25.7	8.2	3.13	Valaoras et al.
						[1988]
Athens 2	Jan – Feb. 1983	TSP / TO	16.1	11	1.46	Valaoras et al.
						[1988]
Paris 1	Fall 1984	TSP / 2ST	4.2	1.6	2.63	Cachier et al.
						[1989]
Paris 2	Winter 1985	TSP / 2ST	14.6	5.9	2.47	Cachier et al.
						[1989]
Gif sur Yvette 1	Winter 1986	TSP / 2ST	6.7	2.4	2.79	Cachier et al.
						[1989]
Gif sur Yvette 2	Spring 1986	TSP / 2ST	6.5	1.9	3.42	Cachier et al.
						[1989]
Gif sur Yvette 3	Summer 1986	TSP / 2ST	4.8	1.5	3.20	Cachier et al.
						[1989]
Gif sur Yvette 4	Fall 1985	TSP / 2ST	18.9	4	4.73	Cachier et al.
						[1989]
Areao	Nov. 1993 – Aug. 1994	0.95 / TO	2.73	0.82	3.33	Pio et al. [1996]
Birmingham,	May 1993	TSP / TO	4.82	1.38	3.49	Castro et al.
						[1999]
Tábua	Jul. 1994 – Aug. 1995	TSP / TO	5.83	1.17	4.98	Castro et al.
						[1999]

Anadia	Aug. 1996	TSP / TO	3.5	1.6	2.19	Castro et al.
						[1999]
Birmingham,	Jan. 1994	TSP / TO	4.78	3.42	1.40	Castro et al.
						[1999]
London 1	Site 1, 1995	TSP / TO	7.6	2.6	2.92	Kendall, et al.
						[2001]
London 2	Site 2, 1995	TSP / TO	6.3	2	3.15	Kendall, et al.
						[2001]
Aspvreten	Jun 1996	10 / 2ST	2.2	0.1	22.00	Zappoli et al.
						[1999]
S. Pietro	Sep. 1996	10 / 2ST	6.2	1	6.20	Zappoli et al.
Capofiume						[1999]
K-Puszta 1	Jul. 1996	10 / 2ST	5	0.6	8.33	Zappoli et al.
						[1999]
K-Puszta 2	Jul. – Aug. 1996	2.5 / RP	10.6	0.42	25.24	Molnár et al.
						[1999]
Coimbra 1	Oct. 1992 – Mar. 1993	TSP / TO	8.88	4.22	2.10	Castro et al.
						[1999]
Coimbra 2	Aug. – Sep. 1993	TSP / TO	5.32	1.78	2.99	Castro et al.
						[1999]
Oporto 1	Apr. – Sep. 1993	TSP / TO	7.22	2.67	2.70	Castro et al.
						[1999]
Oporto 2	Oct. 1992 – Mar. 1993	TSP / TO	9.06	5.33	1.70	Castro et al.
						[1999]
Aveiro	Aug. 1996	TSP / TO	3.04	1.26	2.41	Castro et al.

						[1999]
Basel 1	Winter 1997 – 1998	10 / 2ST O	5.15	3.47	1.48	Röösli, et al.
						[2001]
Basel 2	Spring 1997 – 1998	10 / 2ST O	3.5	2.2	1.59	Röösli, et al.
						[2001]
Basel 3	Summer 1997 – 1998	10 / 2ST O	4.43	2.86	1.55	Röösli, et al.
						[2001]
Basel 4	Fall 1997 – 1998	10 / 2ST O	5.44	4.55	1.20	Röösli, et al.
						[2001]
S. Pietro	Jan. 1998 – Apr. 1999	1.5 / T	8.61	0.53	16.25	Decesari et al.
Capofiume						[2001]
Helsinki	2000 – 2001	2.5 / TO	2.96	1.15	2.73	Viidanoja, et al.
						[2002]
Melpitz	Apr 29 – May 5, 2001	10 / TO	7.7	2.60	2.96	Carvalho et al.
						[2003]
Budapest	23 Apr – 5 May, 2002	2.5 / TO	6.8	3.40	2.00	Salma et al.
						[2004]
Relative standard	deviation (%)		69	85	123	
Relative standard	deviation (%) for BC $> 1.5 \mu g r$	m ⁻³	67	63	36	
North America						
Denver, CO	Nov./Dec. 1978	2.0 / T	7.5	6.4	1.17	Countess et al.
						[1981]
New York, NY	1978/79	3.5 / TO	5.14	3.12	1.65	Shah, et al. [1985]
Detroit, MI 1	July 1981	2.5 / T	7.1	1.6	4.44	Wolff and Korsog
						[1985]

Detroit, MI 2	Jun – July 1981	2.5 / T	5.38	2.32	2.32	Wolff et al. [1985]
Allegheny Mtn.,	Aug. 1983	5.5 / TO	2	1.2	1.67	Keeler et al.
						[1986]
Laurel Hill, MD	Aug. 1983	5.5 / TO	2.3	1.4	1.64	Keeler et al.
						[1986]
Lewes, DE 1	Feb. 1983	2.5 / T	2.4	1.1	2.18	Wolff et al. [1986]
Phoenix, AZ	Jan. 1983	2.8 / 2ST	10	8.3	1.20	Solomon and
						Moyers [1986]
No. Michigan 1	Dec. 1984 – Apr. 1985	TSP / T	2.2	0.48	4.58	Cadle and Dasch
						[1988]
No. Michigan 2	Dec. 1983 – Apr. 1984	TSP / T	1.72	0.72	2.39	Cadle and Dasch
						[1988]
Southern Ontario	July 1986	2.5 / TO	5.2	0.8	6.50	Keeler et al.
						[1990]
Lewes, DE 2	Aug. 1982	2.5 / T	4	0.7	5.71	Wolff et al.
						[1986].
Angeles Natl.	Annual, 1986	10 / TO	6.4	1.2	5.33	Solomon et al.
Forest, CA						[1989]
Detroit, MI 3	Jan. – Apr. 1985	TSP / 2ST	5.7	2.1	2.71	Muhlbaier-Dasch
						[1989]
Detroit, MI 4	Jan. – March 1984	TSP / 2ST	5.5	2.4	2.29	Muhlbaier-Dasch
						[1989]
Crows Landing,	Annual 1988 – 1089	2.5 / TO	3.3	1.46	2.26	<i>Chow et al.</i> [1993]
SJV, CA						
Kern, SJV, CA	Annual 1988 – 1089	2.5 / TO	2.66	1.32	2.02	Chow et al. [1993]

Stockton, CA	Annual 1988 – 1089	2.5 / TO	5.42	3.85	1.41	Chow et al. [1993]
Bakersfield, CA	Annual 1988 – 1089	2.5 / TO	6.5	5.44	1.19	Chow et al. [1993]
Fresno, CA	Annual 1988 – 1089	2.5 / TO	8.05	6.27	1.28	Chow et al. [1993]
Galveston,	Annual 1997 – 1998	2.5 / TO	1.8	0.7	2.57	Fraser et al.
background						[2002]
HRM, supersite,	Annual 1997 – 1998	2.5 / TO	3.3	1.7	1.94	Fraser et al.
TX						[2002]
Bingle, suburban,	Annual 1997 – 1998	2.5 / TO	4	2	2.00	Fraser et al.
TX						[2002]
Clinton, suburban,	Annual 1997 – 1998	2.5 / TO	3.7	2.1	1.76	Fraser et al.
TX						[2002]
Fort Meade, MD	July 1999 – March 2000	2.5 / TO	1.43	1.06	1.35	Chen et al. [2002]
Mexico City 1	Site 1, Feb. – March 1997	2.5 / TO	9.82	4.65	2.11	Chow et al. [2002]
Mexico City 2	Site 2, Feb. – March 1997	2.5 / TO	11	5.59	1.97	Chow et al. [2002]
Mexico City 3	Site 3, Feb. – March 1997	2.5 / TO	12.02	9.39	1.28	Chow et al. [2002]
Mexico City 4	Site 4, Feb. – March 1997	2.5 / TO	9.98	8.28	1.21	Chow et al. [2002]
Mexico City 5	Site 5, Feb. – March 1997	2.5 / TO	7.62	2.89	2.64	Chow et al. [2002]
Mexico City 6	Site 6, Feb. – March 1997	2.5 / TO	8.53	3.75	2.27	Chow et al. [2002]
Look Rock TN 1	Fall 2001	2.5 / TO	3.36	0.55	6.11	Tanner et al.
						[2004]
Lawrence co, TN 1	Winter 2001	2.5 / TO	2.43	0.64	3.80	Tanner et al.
						[2004]
Look Rock 2	Winter 2001	2.5 / TO	1.86	0.65	2.86	Tanner et al.
						[2004]
Look Rock 3	Spring 2001	2.5 / TO	3.57	0.66	5.41	Tanner et al.

						[2004]
Look Rock 4	Summer 2001	2.5 / TO	4	0.66	6.06	Tanner et al.
						[2004]
Lawrence co 2	Summer 2001	2.5 / TO	4.28	0.74	5.78	Tanner et al.
						[2004]
Lawrence co 3	Spring 2001	2.5 / TO	3.57	0.83	4.30	Tanner et al.
						[2004]
Lawrence co 4	Fall 2001	2.5 / TO	4.42	0.91	4.86	Tanner et al.
						[2004]
Chattanooga 1	Summer 2001	2.5 / TO	3.71	1.1	3.37	Tanner et al.
						[2004]
Chattanooga 2	Spring 2001	2.5 / TO	5.71	1.6	3.57	Tanner et al.
						[2004]
Chattanooga 3	Fall 2001	2.5 / TO	7.35	2.2	3.34	Tanner et al.
						[2004]
Chattanooga 4	Winter 2001	2.5 / TO	5.07	2.4	2.11	Tanner et al.
						[2004]
Relative standard deviation (%)			53	94	55	
Relative standard deviation (%) for BC $> 1.5 \mu g \text{ m}^{-3}$			34	60	41	

Table 2. Seasonal changes in OC/BC ratios and BC concentrations (µg m⁻³)

Location	Winter	Spring	Summer	Fall	Average
	OC/BC (BC)	OC/BC (BC)	OC/BC (BC)	OC/BC (BC)	OC/BC
Beijing	2.84 (11.08)	2.73 (6.67)	2.14 (6.27)	2.81 (10.23)	2.63±0.33
a) 1999-00					
Shanghai	2.01 (8.16)	3.05 (5.27)	2.09 (4.61)	2.23 (6.81)	2.34±0.46
b) 1999					
Chongju	1.16 (4.32)	1.34 (3.59)	1.20 (3.37)	0.94 (6.35)	1.16±0.16
c) 1995-96					
Sapporo	0.64 (7.0)	0.82 (3.97)	0.83 (3.03)	1.06 (6.13)	0.84±0.17
d) 1998					
Basel	1.48 (3.47)	1.59 (2.2)	1.55 (2.86)	1.20 (4.55)	1.41±0.18
e) 1997-98					

a) He et al. [2001]; b) Ye et al. [2003]; c) Lee and Kang [2001]; d) Ohta and Okita [1984]; e) Röösli et al. 2001.

Table 3. Comparison of average BC concentrations and apparent and corrected OC/BC ratios

Region	Average BC	Apparent OC/BC	Corrected OC/BC
China	7.10 ± 4.11	2.35 ± 0.61	2.35 ± 0.61
Japan	3.09 ± 5.17	1.29 ± 0.75	1.29 ± 0.75
Korea	3.85 ± 3.39	4.47 ± 4.9	1.47 ± 0.47^x
Europe	2.22 ± 2.09	4.31 ± 5.0	2.41 ± 0.86
			2.27 ± 0.91
North	2.46 ± 2.32	2.92 ± 1.59	2.08 ± 0.85
America			1.88 ± 0.64
All sites	3.36 ± 3.69	3.32 ± 3.48	2.20 ± 1.51
			2.11 ± 1.56
INDOEX	2.71 ± 1.43	3.4 ± 1.7	1.40 ± 0.70
SAFARI		10.50 ± 3.41	6.06 ± 1.85

Table 4. Calculated and corrected measured regional OC/BC ratios

Region	Calculated ^a	Measured
North America		
Contained	1.50	
Biomass	12.69	
Total	4.11	1.88 - 2.08
United States		
Contained	1.52	
Biomass	14	
Total	3.4	1.88 - 2.08
Europe		
Contained	1.22	
Biomass	11.71	
Total	2.40	2.27 - 2.41
China		
Contained	1.54	
Biomass	5.75	

Total	1.89	2.35
India		
Contained	2.82	
Biomass	5.49	
Total	3.25	2.01
Other Asia		
Contained	1.84	
Biomass	7.13	
Total	3.29	Japan 1.29
		Korea 1.47
		Bangladesh 1.84
		INDOEX 1.40

^a From Bond et al., 2004

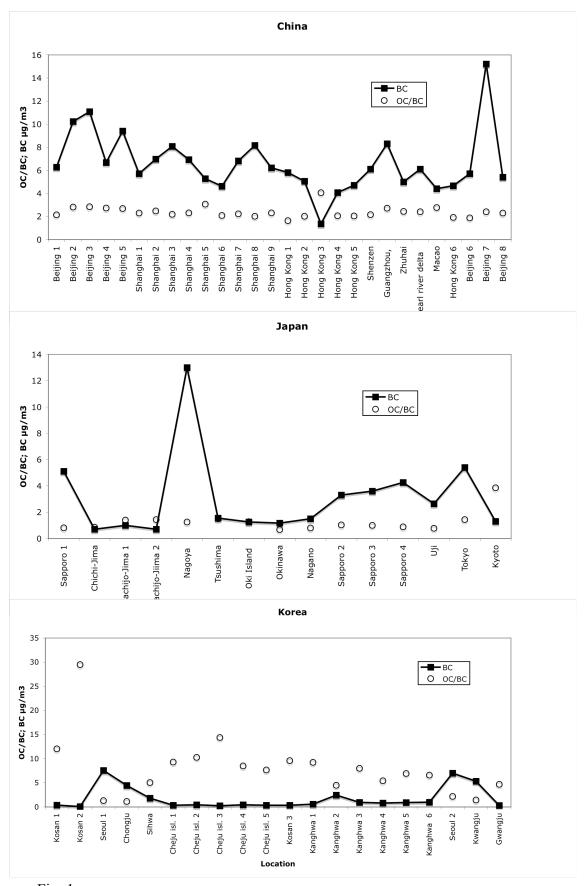
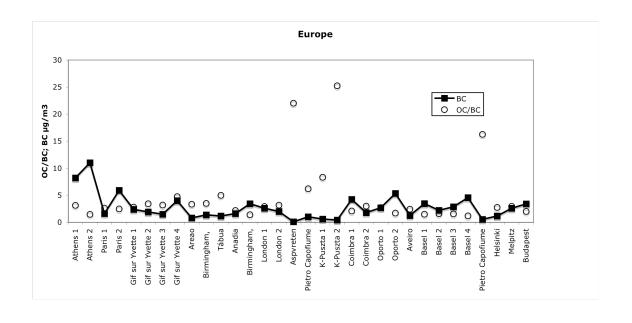


Fig. 1



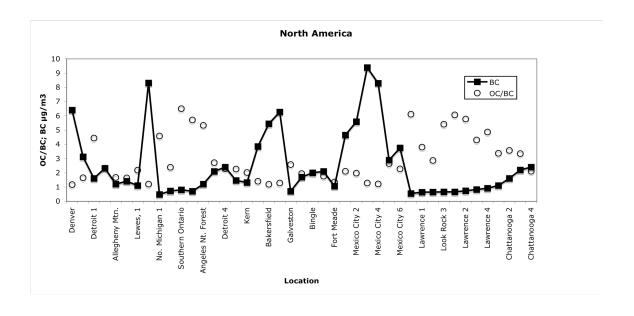


Fig. 2

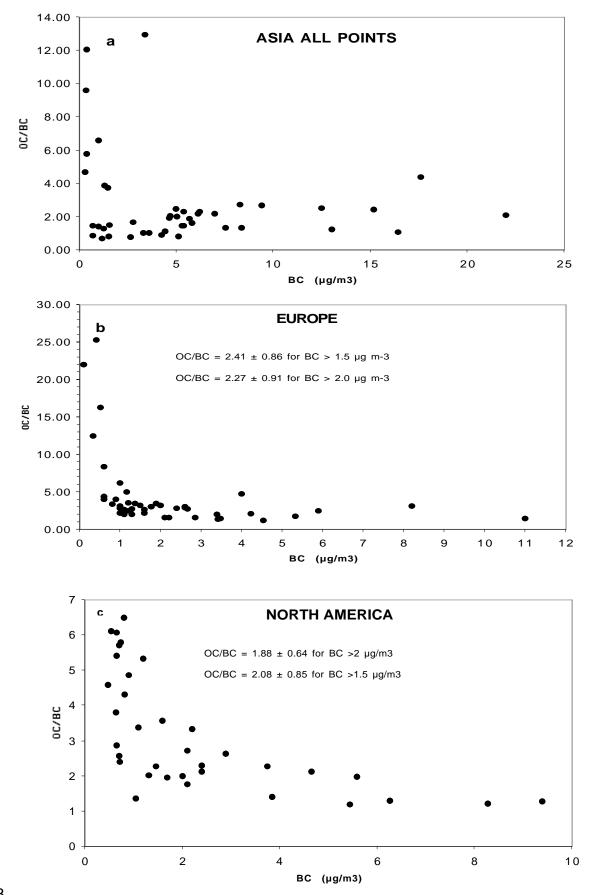


Fig. 3

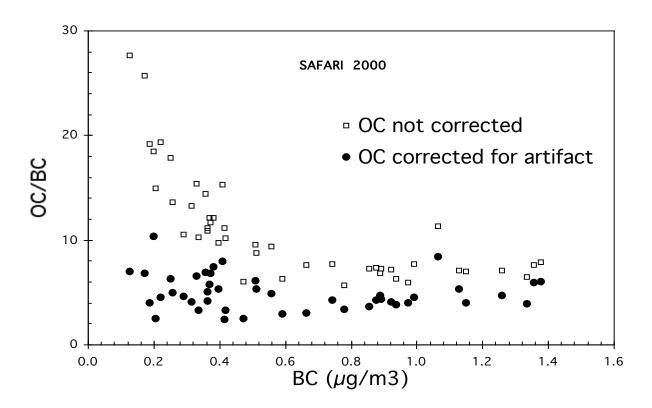


Fig. 4

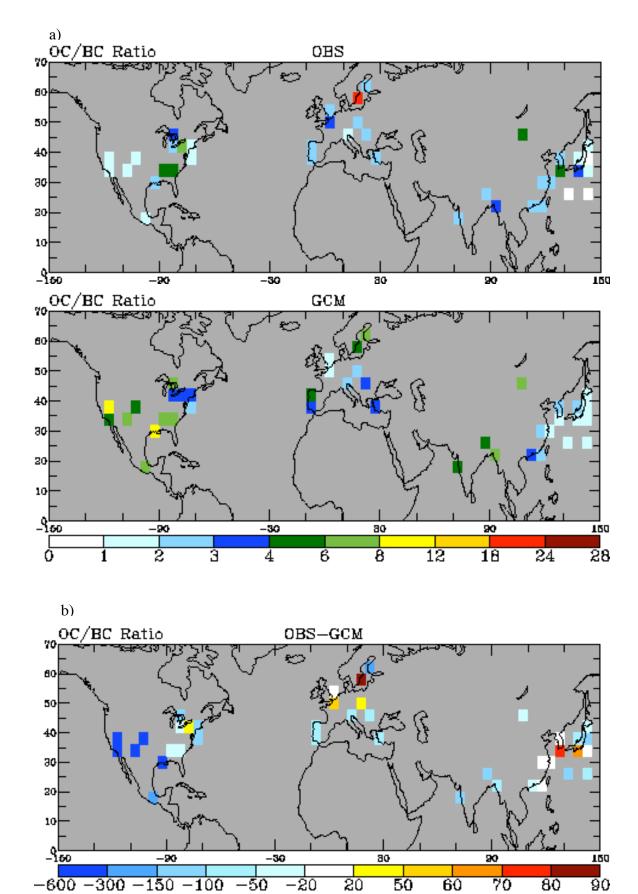


Fig. 5